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Ambient air quality monitoring report for the Waikato Region - 2017

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Executive summary

The main air contaminant of concern in the Waikato Region is PM_{10} (particles in the air 10 microns in diameter or less). Monitoring of PM_{10} was undertaken in the Hamilton, Tokoroa, Taupo, Te Kuiti, Putaruru, Turangi, Morrinsville and Thames airsheds during 2017. In addition to this, PM2.5 (particles in the air 2.5 microns in diameter or less) was monitored in Tokoroa. This particulate monitoring was targeted mainly at identifying the impacts from domestic home heating sources.

The National Environmental Standards for Air Quality (NESAQ) has set a maximum concentration limit for PM₁₀ of 50 μ g/m³ as a 24 hour average. The NESAQ allows for one exceedance of this standard per rolling 12 month period. More than one exceedance within a rolling 12 month period is a breach of the standard and an airshed becomes classified as polluted as a result. A national guideline of 20 μ g/m³ as an annual average for PM₁₀ is also used for assessment purposes. There is currently no New Zealand based guideline or standard applicable to $PM_{2.5}$ and therefore the World Health Organisation (WHO) guideline of 25 μ g/m³ as a 24 hour average and 10 μ g/m³ as an annual average is used for assessment purposes.

In 2017 the 24 hour average standard for PM₁₀ of 50 μ g/m³ was exceeded 10 times in Tokoroa which is the equivalent of nine breaches (based on one allowable exceedance per 12 month period). The 24 hour average standard for PM₁₀ of 50 μ g/m³ was also exceeded once in Taupo and once in Putaruru. There were no exceedances in any of the other airsheds monitored. Annual average concentrations of PM₁₀ were all within the annual average guideline of 20 μ g/m³ in all airsheds.

In Tokoroa, PM_{2.5} concentrations exceeded the WHO guideline of 25 μ g/m³ as a 24 hour average on 53 occasions during 2017. The annual average PM_{2.5} concentration for Tokoroa was 14.7 μ g/m³ which also exceeds the WHO annual average guideline of 10 μ g/m³.

A statistical trends analysis of PM_{10} concentrations indicates a statistically significant decreasing trend in Te Kuiti, Tokoroa, and Putaruru. A previous decreasing trend identified for Taupo is no longer significant and indicates that concentrations are starting to level out. No change has been identified for Hamilton. Improvements have been mainly attributed to reductions in emissions from home heating sources although changes in meteorological conditions have also likely contributed to these observed improvements.

Tokoroa has not complied with its specified NESAQ target of no more than three exceedences of the 24 hour average PM₁₀ standard of 50 μ g/m³ from September 2016. Taupo and Putaruru only just complied with the more stringent NESAQ target of no more than one exceedance of 50 µg/m³ from September 2016. However, ongoing compliance is likely to depend on meteorological conditions and is unlikely to be sustained long term unless further decreases in emissions occur. Te Kuiti now meets the criteria of "not polluted" having been compliant with the NESAQ for over five years which is a pre-requisite for achieving this status under the NESAQ.

While woodburners for domestic home heating have been identified as the main source of poor air quality in all airsheds, traffic sources in Hamilton have been identified as significant contributors to poor air quality in localised areas close to busy traffic routes and intersections based on the results of nitrogen dioxide $(NO₂)$ monitoring.

Monitoring of NO² was undertaken by NZTA in Hamilton, Cambridge, Te Awamutu and Taupo, in 2017 for purposes of identifying the impacts from traffic related emissions along state highways. Monitoring of benzene, toluene, xylenes and ethylbenzene (BTEX) were also undertaken in Hamilton by WRC in 2017 as additional indicators of traffic related emissions.

The results from NO₂ monitoring indicate that the Lorne Street/Ohaupo Road site and Greenwood Road/Killarney Road site in Hamilton continue to exceed the WHO annual guideline

of 40 μ g/m³ for NO₂. No exceedances of the WHO annual guideline for NO₂ at the other Hamilton sites and the Cambridge, Te Awamutu and Taupo sites have been identified. Long term seasonal trend analysis provides no indication of either a worsening or improving trend for monthly average NO₂ concentrations at any of the monitoring sites for the period 2011 to 2017.

Concentrations of benzene in Hamilton were within the national guideline of 3.6 μ g/m³ at all sites and show evidence of a decreasing trend. The highest annual concentration was measured at the Greenwood Street monitoring site and was 2.2 μ g/m³. Concentrations of toluene, ethylbenzene and xylene were also well within acceptable international criteria.

1 Introduction

1.1 Background

New Zealand (and the Waikato region) enjoys good air quality most of the time due to the country's long narrow shape, exposure to strong eastward winds, and typically low population density. However, during the winter months when temperature inversions form on cold, still days, air pollutants can become trapped close to the ground leading to poor air quality in some towns and cities.

Air quality within towns and cities can be affected by a number of activities including natural causes such as sea spray, pollen and volcanic activity and human created causes such as home heating, traffic and industrial discharges.

Although air pollution is a complex mixture of contaminants and particles, the majority of health effects in New Zealand are associated with particulate matter less than 10 micrometres in size – commonly known as PM10. It is also a good indicator of the sources and effects of other air pollutants. However, while PM_{10} levels pose a risk to human health, it is now well established internationally, that the finer particle range, referred to as $PM_{2.5}$, provides better evidence of effects on human health and is more indicative of the problem source, which is combustion related.

The Health and Air Pollution in New Zealand Study (Kuschel et al., 2012) identified that domestic fires (open fires and wood and coal burners used for home heating) dominate the health impacts associated with PM₁₀ in every location in New Zealand except Auckland.

Emissions from motor vehicles are also sources of air pollutants that have adverse health effects. Exhaust emissions include fine particulate matter, volatile organic compounds (such as benzene) and the gases carbon monoxide and nitrogen oxides. Nitrogen dioxide emissions can occur directly from combustion processes and as a result of the conversion of nitric oxide gas (also produced from combustion processes) reacting in the atmosphere in the presence of ozone. In New Zealand, motor vehicle emissions are the main source of nitrogen dioxide in urban areas.

Other gases (such as ozone and sulphur dioxide) and secondary particulate (sulphates and nitrates) can form in the atmosphere from reactions involving some of these primary emissions. Sulphur dioxide is typically associated with combustion of fuels containing high levels of sulphur such as coal from industry and heavy fuel oils used in shipping.

Ambient air quality monitoring has been carried out in the Waikato region by the Waikato Regional Council (WRC) since 1998. The US EPA define ambient air monitoring as the systematic, long-term assessment of pollutant levels by measuring the quantity and types of certain pollutants in the surrounding, outdoor air.

1.2 Regulatory requirements and assessment criteria

In New Zealand, monitoring and management of air quality is undertaken within defined air quality management areas referred to as airsheds. Airshed boundaries are officially identified by regional councils and approved and made public by the Ministry for the Environment (MfE).

Regional councils have a responsibility to monitor and manage outdoor air quality under the Resource Management Act 1991 (RMA). The National Environmental Standards for Air Quality (NESAQ) are mandatory environmental regulations made under the RMA that:

- \bullet direct regional councils to focus on monitoring of PM₁₀ as the main contaminant of concern and on managing the main source of PM_{10} emissions, namely woodburners used for domestic home heating;
- include ambient air quality standards for PM_{10} , carbon monoxide, nitrogen dioxide, sulphur dioxide and ozone for protecting human health;
- require regional councils to monitor air quality if it is likely that the ambient air quality standard for a contaminant will be breached in an airshed;
- state that an airshed is classified as polluted if it has more than one PM_{10} exceedance per 12 month period; and
- state that an airshed ceases to be polluted when the PM_{10} standard has not been breached in the airshed for five years.

The air quality standard for PM₁₀ is 50 micrograms per cubic metre of air (50 μ g/m³) averaged over a 24-hour period (from midnight to midnight), with one allowable exceedance per 12-month period. The standards for PM10, carbon monoxide, nitrogen dioxide, sulphur dioxide and ozone along with averaging periods and number of allowable exceedances are provided in Table 1.1. A breach of the standards occurs when more than the allowable number of exceedances occurs within the specified period.

Table 1.1: National Environmental Standards for Air Quality.

a. Note that the 1-hour average concentration of sulphur dioxide can exceed 350 μ g m⁻³ up to 9 times within any 12-month period but can never exceed a 1-hour average of 570 μ g m $^{-3}$.

For the assessment of other air contaminants not included under the NESAQ, the National Ambient Air Quality guidelines (MfE, 2002) and the World Health Organisation guidelines (WHO, 2005) are used. A subset of ambient air quality guidelines relevant to WRC's air quality monitoring programme are provided in Table 1.2.

Table 1.2: Ambient air quality guidelines.

There are 20 airsheds in the Waikato region that have been officially gazetted by MfE for monitoring and management purposes(refer to Figure 1.1). To date, a total of 14 of these airsheds have been monitored for PM₁₀. A combination of long-term monitoring and short-term surveys are undertaken as described below:

• Long-term monitoring is undertaken for airsheds that are shown to either breach or have the potential to breach the ambient air quality standards. There are currently three polluted airsheds in the Waikato region, namely Putaruru, Taupo, and Tokoroa. Te Kuiti and Hamilton, although not currently classified as polluted, are also included for longterm monitoring purposes due to previous breaches and or ongoing potential for future breaches.

• A three-yearly "rolling" survey is undertaken for airsheds that have not yet been monitored. In these cases, WRC's monitoring stations are moved to new locations according to a three year rotation subject to the airshed having had no exceedances of the 24 hour average PM_{10} standard over that three year period.

Figure 1.1: Map of Waikato Region indicating the 20 urban centres that have been gazetted as airsheds.

The NESAQ also sets PM_{10} compliance targets for polluted airsheds. Polluted airsheds which have had less than 10 exceedances per year, must not breach the standard after September 2016 (i.e. Taupo, and Putaruru). Airsheds that have had 10 or more exceedances per year, must have no more than three exceedances after September 2016 and no more than one exceedance after September 2020 (i.e. Tokoroa). As of December 2016, the Te Kuiti airshed is now classified as nonpolluted as it has not had a breach of the NESAQ for five years. The Hamilton airshed is also not classified as polluted as it has not had any breaches of the standard since 2009 apart from a series of breaches in 2013 which were considered exceptional circumstances and were discounted by $MfE¹$.

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¹ An application was made to the Minister for the Environment in August 2013 and an exemption for the eight exceedances was provided by the Minister in November 2013. As a result, the exceedances do not count towards determining the polluted status

1.3 Monitoring objectives

The objectives of Waikato Regional Council's ambient air quality monitoring programme are to:

- 1. Determine compliance with national ambient air quality standards and guidelines.
- 2. Identify trends in ambient air quality and sources contributing to poor air quality.
- 3. Provide information to support the implementation of our regional plan and regional policies on air quality, including consent processing and consented and permitted activity compliance.

1.4 Airsheds and contaminants monitored in 2017

Ambient air quality and meteorological monitoring for 2017 was undertaken in the following airsheds:

- Hamilton (PM_{10} , $PM_{2.5}$, NO₂, benzene, ethyl benzene, toluene, xylenes and meteorology)
- Taupo (PM_{10} , NO₂ and meteorology)
- Tokoroa (PM₁₀, PM_{2.5}, NO₂ and meteorology)
- Te Kuiti (PM_{10} and meteorology)
- \bullet Putaruru (PM₁₀)
- \bullet Turangi (PM₁₀)
- Cambridge $(NO₂)$
- Te Awamutu-Kihikihi $(NO₂)$
- Morrinsville (PM_{10})
- Thames (PM_{10})

Refer to Figures 1.2 to 1.5 for maps of the airsheds monitored in 2017. All stations, except Hamilton's Bloodbank station are set up for assessing ambient air quality including the impacts of residential solid fuel heating. The Bloodbank station is located 10 metres away from a busy roadway (Ohaupo Road) and is near to industry discharges (e.g. Waikato Hospital boilers) and is therefore expected to be more influenced by traffic and industry sources compared with the other Hamilton station at Claudelands.

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of the Hamilton airshed i.e. Hamilton remains a non-polluted airshed and the industry offset and open fire place ban regulations of the NESAQ are not triggered.

Figure 1.2: Hamilton, Tokoroa and Taupo 1

Figure 1.3: Te Kuiti, Putaruru and Turangi airsheds with monitoring stations identified by pink triangle.

Figure 1.4: Cambridge, Te Awamutu-Kihikihi and Morrinsville airsheds with monitoring stations identified by pink triangle.

Figure 1.5: Thames airshed with monitoring station identified by pink triangle.

Hamilton is the Waikato region's main urban area, with a population of 141,612. The airshed extends across both the eastern and western sides of the Waikato River. The main source of poor air quality during winter is from home heating using wood. However, traffic can also be a significant source of poor air quality especially in the vicinity of busy intersections. Hamilton currently has two active ambient PM_{10} monitoring stations and six localised traffic BTEX monitoring sites. The main ambient air quality monitoring station is at the Claudelands Event Centre on Heaphy Terrace on the eastern side of Hamilton. The Bloodbank station is a secondary air quality monitoring station, more influenced by localised traffic and industry sources. It is next to the NZ Blood Service at the Waikato Hospital on the corner of Ohaupo Road and Lorne Street on the western side of Hamilton. From November 1997 to October 2013, the main ambient air quality monitoring station waslocated on Peachgrove Road on the eastern side of Hamilton. The station was decommissioned in October 2013 because the property changed ownership.

Tokoroa is located mid-way between Hamilton and Taupo on State Highway One, close to the foot of the Mamaku Ranges with a population of 12,243. The main source of poor air quality during winter is home heating using wood. Tokoroa's one active ambient air quality and meteorological monitoring station is located at the Billah Street Reserve and was established in March 2001. In April 2015, the station was shifted (within the same site) to a new enclosure 25 metres from its previous location because of concern about potential impacts from the lime dosing process used on site by South Waikato District Council for treating drinking water at the Billah Street reservoir.

Taupō, with a population of around 23,700, is the Waikato region's second largest urban area after Hamilton. It is located within the Taupō Volcanic Zone at the northeastern end of Lake Taupō. The main source of poor air quality during winter is from home heating using wood. Because of its location within the Taupō Volcanic zone, there will be a larger contribution from volcanic/geothermal activity at times. The surrounding pumice soils are also likely to contribute to a higher summertime dust source. Taupō's one active ambient air quality and meteorological monitoring station, located at Gillies Avenue Reserve, was established in November 2000.

Te Kuiti is located approximately 78km south of Hamilton within a confined valley that is aligned along an approximately southeasterly to northwesterly direction. The town is located within the Waitomo District and has a population of 4221. The main source of poor air quality during winter is home heating using wood but there are also three significant industrial sources including a lime processing plant and two woodfired combustion plants associated with two separate sawmilling operations. Te Kuiti's ambient air quality monitoring and meteorological station was established at the Waitomo District Council offices on Queen Street in May 2003.

Putaruru is located 65 kilometres southeast of Hamilton and is close to Lake Arapuni on the Waikato River. It is situated midway between Tokoroa and Tirau on State Highway One, in the

South Waikato District and has a population of 3777. Putaruru occupies a flat to gently undulating site, and to the east the land rises to the Mamaku Range. The main source of poor air quality is home heating. Putaruru's ambient air quality monitoring station was established at the Bowling Club on Arapuni Street, in July 2006.

Turangi, at the southern end of Lake Taupō, has a population of 2952. The main source of poor air quality is home heating. Due to its location within the Taupo Volcanic zone, at times volcanic/geothermal activity has been identified as contributing to PM_{10} concentrations. Turangi's ambient air quality monitoring station was established at the Turangi Fire Station on Ohuanga Road in March 2009.

Cambridge, approximately 23km southeast of Hamilton, has a population of approximately 18,000. The airshed extends across both the eastern and western sides of the Waikato River. The main source of poor air quality is home heating. PM_{10} was monitored in Cambridge at an ambient air quality monitoring station located at Leamington Domain on Scott Street from May 2013 to August 2016. NZTA have been monitoring $NO₂$ at the intersection of Victoria St and Queen St since 2007.

Te Awamutu, approximately 30km south of Hamilton, has a population of 10,305. The main source of poor air quality is likely to be home heating but there are also significant industrial emissions associated with the Fonterra dairy factory which has milk powder driers and gas and coal fired boilers. PM₁₀ was monitored in Te Awamutu at an ambient air quality monitoring station located at Albert Park on Albert Park Drive from June 2013 to October 2016. NZTA have been monitoring NO² at the intersection of Ohaupo Road and Albert Drive since 2010.

Morrinsville, approximately 30km northeast of Hamilton, has a population of approximately 7000. The main source of poor air quality is home heating. Morrinsville's ambient air quality monitoring station was established at Morrinsville College in the vicinity of North Street in May 2015.

Thames is located at the southeastern end of the Firth of Thames on the banks of the Kauaeranga River. Most of the urban area occupies a coastal strip of flat land at the western base of the Coromandel Range. Thames has a population of 7,518. The main source of poor air quality is home heating. Thame's ambient air quality monitoring station was established at Thames High School in the vicinity of Richmond Street in March 2016.

2 Methodology

2.1 PM¹⁰ & PM2.5 monitoring

There are many different ways of measuring particle concentrations (e.g. PM_{10} or $PM_{2.5}$ etc.) in air and each method can provide varying levels of accuracy. One of the most common methods used by Regional Councils around New Zealand is the BAM. BAM stands for Beta Attenuation Monitor. This instrument type measures particle concentrations continuously. Sampled air is drawn through a filter which collects particles. A beam of beta radiation (electrons) is passed though the filter and counted on the other side of the filter by a detector. As the mass of particles on the filter increases the beta count is reduced and from this the mass of particles can be determined. This method is a US EPA equivalent method which is therefore accepted for compliance purposes under the NESAQ.

All air quality monitoring stations in the Waikato region used a BAM for monitoring particulate concentrations in 2017.

Refer to table 2.1 for details of monitoring station locations and instruments used for 2017. All stations, except Hamilton's Bloodbank station are set up for assessing the impacts of residential solid fuel heating. The Bloodbank station is located 10 metres away from a busy roadway (Ohaupo Road) and is near to industry discharges (e.g. Waikato Hospital boilers) and is therefore expected to be more influenced by traffic and industry sources compared with the other Hamilton station at Claudelands.

Table 2.1: PM¹⁰ and PM2.5 monitoring stations and instruments for 2017.

1. Note that prior to 2014, the main monitoring station in the Hamilton airshed was located 160 Peachgrove Road from 1997 to 2013. The Claudelands site is a replacement for this earlier "Peachgrove Rd" site.

 PM_{10} and PM_{2.5} data from the BAM monitors are recorded and logged by iQuest iRIS 320 dataloggers and telemetered hourly to Waikato Regional Council and stored in the hydrotel database. Subsequent data processing and archiving is undertaken in the WISKI database. Over the period 2006 to 2010, several air quality monitoring stations in the Waikato region had Thermo Sequential Partisol gravimetric samplers run in conjunction with existing BAM instruments in order to determine the need to apply site-specific corrections to the BAM PM_{10} data. As a result, it was determined that site-specific corrections were required for BAM PM₁₀ data collected at Tokoroa, Taupo and Putaruru. Monitoring data has therefore been adjusted accordingly, at those sites. Table 2.2 provides details of the formulae used for applying the corrections. Additional assessments using a sequential partisol have been undertaken in Tokoroa in 2014 and 2016 and indicate that the currently applied correction factor for BAM PM₁₀ data for Tokoroa is still valid. No corrections are applied to the $PM_{2.5}$ data.

Table 2.2:	Site-specific corrections applied to BAM PM_{10} data.		
	Station	Correction factor	
	Tokoroa	Corrected $PM_{10} = 10^{(1.09945\log BAM - 0.08595)}$	
	Taupo	Corrected $PM_{10} = 1.255BAM - 1.538$	
	Putaruru	Corrected $PM_{10} = 1.106$ BAM - 2.38	

Table 2.2: Site-specific corrections applied to BAM PM¹⁰ data.

2.2 BTEX monitoring

Passive sampling for the volatile organic compounds benzene, ethyl-benzene, toluene and xylenes (BTEX) was undertaken in Hamilton at six sites in 2017 using 3M Passive axial diffusion samplers. Two of the six sites are located on Peachgrove Road adjacent to Countdown supermarket and Hamilton Intermediate School. The other four sites are located at Bridge Street/Grey Street intersection, Claudelands Road/Victoria Street intersection, Tristram Street and Greenwood Street (refer to Figure 2.1).

Motor vehicles are the main sources of BTEX in urban areas. Pollutants such as benzene are typically monitored using a passive sampling methodology where small plastic badges containing activated carbon are attached to a power pole, for example, in proximity to a busy street or intersection and are exposed for a certain period to the ambient air. The benzene is adsorbed to the carbon over this period and can then be analysed in a laboratory at the end of the exposure period.

Figure 2.1: Map of BTEX sites in Hamilton 1

The method used is as described in Stevenson and Narsey (1999) with passive axial samplers being deployed for periods of three months. The BTEX compounds are extracted from the samplers using carbon disulphide and then analysed using Gas Chromatography Mass Spectrometry by Hill Laboratories in Hamilton. While this type of passive sampling is recommended as a screening method only, it is the most common approach to BTEX monitoring in New Zealand and is significantly more cost effective than the radial passive sampler method recommended by the Ministry for the Environment's ambient air quality guidelines (MfE 2002).

Validation of this axial sampler method with the MfE recommended method was undertaken in 2010 by Waikato Regional Council. It was determined that the accuracy of the 3M passive axial samplers is likely to be sufficient for monitoring purposes provided the deployment duration is not excessive (Mathieson, 2010).

The 2017 monitoring period for BTEX covered the period 4 January to 6 December 2017. Aligning the reporting period for BTEX with an exact calendar year is not always feasible as it would require staff to deploy passive sampling equipment on 1 January. This misalignment with the calendar year is not of concern when reporting annual averages as it is unlikely that there would be any difference in the annual averages if the passive samplers were instead deployed on 1 January each year.

2.3 Nitrogen dioxide monitoring

Nitrogen dioxide ($NO₂$) monitoring is used as a general proxy for air pollution from motor vehicles. The NZ Transport Agency undertakes passive NO₂ monitoring near to state highways across New Zealand. The Waikato Regional Council partners with NZTA to collect additional NO₂ data at selected sites in the Waikato region including Hamilton, Te Awamutu, Taupo and Cambridge. Monitoring was undertaken in these four airsheds in 2017 for a full 12 month period. The locations of monitoring sites are provided in Table 2.3 and Figures 2.2 to 2.3.

Table 2.3: Site locations for passive NO² monitoring in 2017.

Passive nitrogen dioxide samplers (similar to the BTEX samplers described above) are easy to operate and relatively cheap, so they can be installed in large numbers (usually attached to power poles for example) over a wide area giving good spatial coverage. The sampler tubes contain triethanolamine which reacts with nitrogen dioxide to form nitrite ions. The tubes are typically exposed for a one month duration and then analysed at a laboratory using UV spectroscopy.

The results are indicative only and provide monthly rather than daily averages. Passive sampling is therefore useful as a screening method rather than a regulatory method, for which continuous monitors are used.²

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² Ambient air quality (nitrogen dioxide) monitoring network report 2007 - 2009, NZTA

Figure 2.2: M**ap of passive NO² monitoring sites in Hamilton.**

Figure 2.3: Map of passive NO² monitoring sites in Cambridge. Te Awamutu and Taupo.

2.4 Meteorological monitoring

In 2017, meteorological data, including ambient air temperature, wind speed and wind direction were collected at Hamilton (Claudelands), Tokoroa, Taupo, Matamata and Te Kuiti monitoring stations. Relative humidity data was also collected at Hamilton (Claudelands) and Taupo. All meteorological data is collected at a height of six metres above ground level except for Tokoroa where meteorological data is collected from the top of an existing water reservoir tower at a height of 10 metres. Details of the meteorological instrumentation used is provided in Table 2.4.

The meteorological data is recorded and logged by iQuest iRIS 320 dataloggers and telemetered hourly to Waikato Regional Council and stored in the hydrotel database. Subsequent data processing and archiving is undertaken in the WISKI (by Kisters) database.

The frequency and extent of NES breaches from year to year depends largely on the prevalence of meteorological conditions conducive to elevated pollution, in particular low wind speeds, cooler temperatures and temperature inversions.

A graphical comparison of hourly average PM_{10} concentrations on days where the 24-hour average is elevated or exceeds the standard relative to the hourly average wind direction, wind speed and air temperature provides a means of identifying contributing meteorological influences. This type of comparative assessment has been applied to PM_{10} and meteorological data for Hamilton, Taupo, Te Kuiti and Tokoroa for 2017.

A wind rose is a diagram that shows the relative frequency and speed of winds from different directions at a specific location. They provide a useful means of visually identifying prevailing wind directions and windspeed frequencies. The application, WRPLOT View (Version 7.0.0) by Lakes Environmental, has been used to generate wind roses for Hamilton, Tokoroa, Taupo and Te Kuiti.

2.5 Trend analysis

Analysis of trends in PM¹⁰ concentrations requires an assessment of the variability from year to year occurring because of meteorological conditions. For example, higher PM10 and more exceedances might be expected if a winter has a greater number of days when the wind speed is low and there is vertical stability in the lower atmosphere. Over the short term this interannual variability will mask any genuine underlying trend toward better or worse air quality. A reasonably long monitoring record is therefore needed to confirm or exclude the possibility of any underlying trend.

Seasonal Mann-Kendall test for monotonic trends (consistently increasing or decreasing trends) can be used for detecting underlying trends in variable environmental time-series data sets, and may suggest presence of an underlying trend which is not evident from visual inspection of the PM_{10} record or summary statistics. This method generates probability (p) values that are used to assess the likelihood that the apparent relationship is real or a result of chance. The conventional threshold for deciding whether a relationship is likely to be genuine is at a probability value of p = 0.05 or lower, which corresponds to a 95% confidence level and greater. A negative MK-Stat indicates a decreasing trend.

An updated Seasonal Mann-Kendall trend analysis was undertaken for Hamilton, Tokoroa, Taupo, Te Kuiti and Putaruru for PM_{10} datasets up to and including 2017. This analysis was also undertaken for NO₂ datasets collected by NZTA for Cambridge, Hamilton, Taupo and Te Awamutu.

3 Results

3.1 PM¹⁰ monitoring in Hamilton

The Hamilton airshed remains compliant with the NES with no exceedances of the PM₁₀ standard of 50 μ g/m³ as a 24-hour average. The maximum 24-hour average PM₁₀ concentration measured was 40 μ g/m³ at the Ohaupo Road (Bloodbank) site on 9 June 2017 and 30 μ g/m³ at the Claudelands site on 12 October 2017.

Figure 3.1: 24-hour average PM¹⁰ concentrations measured at Ohaupo Road during 2017.

Figure 3.2: 24-hour average PM¹⁰ concentrations measured at Claudelands during 2016.

Figure 3.3 shows year to year variability in 24-hour average PM $_{10}$ concentrations relative to MfE air quality indicator categories from 2014 to 2017 for the Claudelands site. The proportion of 24-hour average PM₁₀ concentrations falling within the good air quality category (less than 33% of the guideline) for 2017 at 89% indicates a small improvement compared with previous years.

Figure 3.3: Comparison of 24-hour average PM¹⁰ concentrations measured at Claudelands in Hamilton from 2014 to 2017 relative to air quality indicator categories.

A comparison of PM₁₀ concentrations at the Ohaupo Rd site from 2012 to 2017 relative to the MfE air quality indicator categories are shown in Figure 3.4. The comparison indicates that there has been a gradual increase in the data falling within the good air quality category (less than 33% of the air quality guideline) with 84% of data falling within this category for 2017 compared with only 74% of data in 2012 and 2014.

Figure 3.4: Comparison of 24-hour average PM¹⁰ concentrations measured at Ohaupo Road site in Hamilton from 2012 to 2016 relative to air quality indicator categories.

A comparison of monthly average PM₁₀ concentrations at Ohaupo Road and Claudelands for 2017 (Figure 3.5) indicates slightly higher overall concentrations at the Ohaupo Road site but with a similar seasonal distribution. Interestingly, the seasonal distribution pattern indicates a similar peak in January compared with June which suggests a summertime/ non-homeheating impact on PM_{10} concentrations that needs further investigation. This summer and winter seasonal pattern has been observed in previous years for Hamilton and is inconsistent with the more pronounced wintertime peak observed for other airsheds in the Waikato region.

Figure 3.5 Comparison of monthly average PM¹⁰ concentrations measured at Claudelands and Ohaupo Rd sites in Hamilton in 2017.

The annual average PM₁₀ concentrations for Ohaupo Road and Claudelands for 2017 were 12.1 μ g/m³ and 11.0 μ g/m³ respectively which is well below the annual average guideline for PM₁₀ of 20 μ g/m³ (MfE, 2002). An annual average PM₁₀ concentration is not specified in the NES. Refer to Appendix 1 for a full summary of PM₁₀ monitoring statistics for Hamilton for the period 2002 to 2017.

A windrose showing the relative frequency and speed of winds from different directions for 2017, as measured at the Claudelands station, is compared with the windrose for 2016 in Figure 3.6. The 2017 wind rose indicates a prevailing wind direction from the west and west-southwest with a high frequency of low to moderate windspeeds which is very consistent with the windrose pattern observed for 2016.

Figure 3.6: Windrose of wind direction and windspeed data as measured at Claudelands station in Hamilton in 2016 and 2017.

A comparison of meteorological parameters with hourly average variations in PM_{10} on the three days during 2017 which had the highest 24-hour average PM_{10} concentrations, all recorded at the Ohaupo Road site, is provided in Figure 3.7. These days were 9 June, 27 June and 7 September (PM₁₀ = 39.7 μ g/m³, 33.4 μ g/m³ and 33.1 μ g/m³ respectively).

The variation in PM_{10} concentrations throughout the 27 June is more typical of the daily concentration pattern (i.e. higher concentrations in the evening with a smaller peak in the morning) observed in urban areas of New Zealand primarily because of solid fuel burning for domestic heating (Trompetter *et al.*, 2010). In comparison the PM₁₀ concentration on 9 June peaked significantly in the morning compared with a much smaller peak in the evening. The air temperatures and wind speeds were similar on both the 9 and 27 June with wind speeds of 2 m/s or less throughout the day and air temperatures significantly lower in the morning.

The variation in PM_{10} concentrations on 7 September indicates no morning peak but a large peak in the evening and a smaller one in the afternoon. Wind speeds were much higher throughout the day and air temperature did not drop significantly in the morning compared with the evening.

Figure 3.7: Hourly average PM¹⁰ concentrations, wind speed, wind direction and temperature on days when PM¹⁰ concentrations were elevated in Hamilton.

Seasonal Mann Kendall test results for monthly average PM_{10} concentrations for the Ohaupo Road site for the period 2012 to 2017 indicate no statistically evident change (MK-Stat of -1.2 and p-value of 0.21). Test results for monthly average PM_{10} concentrations for the Claudelands site for the period 2014 to 2017 also indicate no statistically evident change (MK-Stat of -1.1 and p-value of 0.27).

Data suggests that Hamilton is likely to continue to meet the NES target of no more than one exceedance per 12-month period of the 24-hour average PM_{10} concentration standard of 50 µg/m³. However, it is possible that poor meteorological conditions i.e., a particularly cold and calm winter could still result in more than one exceedance within a 12-month period. This would result in Hamilton being classified as a polluted airshed.

3.2 BTEX monitoring in Hamilton

Benzene concentrations measured at all locations in Hamilton during 2017 were lower than the Ministry for the Environment's annual average guideline of 3.6 μ g/m³ (refer to Figure 3.8). The guideline prior to 2010 was 10 μ g/m³ (annual average). The highest annual average benzene concentration of 1.9 μ g/m³ was measured at the Greenwood Street monitoring site. An improving or "levelling" trend is evident for annual average concentrations of benzene at all sites with the exception of Peachgrove Road which shows an increase for 2016 and 2017 compared with the previous 3 to 4 years. A new supermarket opening on this road, near to the monitoring site location at the end of 2015 is the likely cause of the increase observed.

Large decreases in benzene concentrations were observed over the period 2003 to 2007. These earlier decreases were attributed to changes in fuel specifications and improved vehicle technology (Smith, 2007).

Figure 3.8: Annual average benzene measured at Hamilton sites (2003 to 2017).

Ethyl-benzene, toluene and xylene were also measured at the BTEX monitoring sites for each year. Results are shown in Table 3.1. An MfE document discussing amendments to the 1994 ambient air quality guidelines suggests an annual threshold of 190 μ g/m³ and 950 μ g/m³ for toluene and total xylenes respectively (MfE, 2000). The United States EPA Reference Concentration for ethyl-benzene is 1000 μ g/m³ as an annual average. Concentrations of toluene, total xylenes and ethyl-benzene measured in Hamilton at all sites were significantly lower than the suggested MfE thresholds and USEPA Reference Concentration.

a There are currently no guideline values for ethyl-benzene, toluene and xylenes. Threshold values for toluene and total xylenes used here are from proposed amendments to the 1994 ambient air quality guidelines. For ethyl-benzene, the US EPA Reference Concentration has been used.

3.3 NO² monitoring in Hamilton

The results from NZTA's passive $NO₂$ monitoring programme for 2017 (Figure 3.9) indicate that the Lorne Street/Ohaupo Road site and Greenwood Road/Killarney Road site in Hamilton have some of the highest readings in New Zealand. Both sites have exceeded the World Health Organisation (WHO) annual guideline of 40 μ g/m³ for NO₂. Seasonal Mann Kendall test results for monthly average $NO₂$ concentrations for the Hamilton sites for the period 2011 to 2017 indicate no statistically evident change.

Figure 3.9: Annual average NO² measured at Hamilton sites (2007 to 2017).

3.4 PM¹⁰ and PM2.5 monitoring in Tokoroa

Concentrations of PM₁₀ in Tokoroa exceeded the PM₁₀ standard of 50 μ g/m³ as a 24-hour average on ten occasions during 2017 (Table 3.2). The maximum 24-hour average PM_{10} concentration measured was 75 μ g/m³ on 16 June 2017 and is higher than the maximum 24-hour average concentration of 60 μ g/m³ recorded in 2016.

hour average in Tokoroa during 2017.				
Date	$PM_{10} \mu g/m^3$	Rank		
26 June	$51 \mu g/m^3$	$\mathbf{1}$		
16 July	51 μ g/m ³	$\overline{2}$		
7 June	52 μ g/m ³	3		
11 July	54 μ g/m ³	4		
31 July	56 μ g/m ³	5		
30 July	57 μ g/m ³	6		
27 June	58 μ g/m ³	$\overline{7}$		
8 June	59 μ g/m ³	8		
9 June	63 μ g/m ³	9		
16 June	75 μ g/m ³	10		

Table 3.2: Dates and concentrations for exceedences of the PM¹⁰ standard of 50 µg/m³as a 24 hour average in Tokoroa during 2017.

During 2017 PM_{2.5} concentrations in Tokoroa exceeded the World Health Organisation (WHO) guideline of 25 μ g/m³ (24-hour average) on 53 occasions. The maximum 24 hour average PM_{2.5} concentration of 65 μ g/m³ was measured on 30 July 2017 which is similar to the maximum of 66 μ g/m³ recorded for PM_{2.5} in 2016.

Daily PM₁₀ and PM_{2.5} concentrations measured at Tokoroa during 2017 are shown in Figures 3.10 and 3.11. Highest PM₁₀ and PM_{2.5} concentrations were measured during the winter months when emissions from solid fuel burning coincides with meteorological conditions conducive to elevated concentrations.

Figure 3.10: 24 hour average PM¹⁰ concentrations measured in Tokoroa during 2017.

Figure 3.11: 24 hour average PM2.5 concentrations measured in Tokoroa during 2017.

Figure 3.12 shows the relationship between PM_{10} and $PM_{2.5}$ concentrations for Tokoroa for winter 2017. A good correlation between $PM_{2.5}$ and PM_{10} is observed (r^2 =0.85) for 2017 as was observed for 2015 and 2016. The slope is also less than one compared with that observed for 2016 indicating that generally PM_{2.5} concentrations were less than PM₁₀. This is more consistent with the expectation that as $PM_{2.5}$ is a subset of PM_{10} the $PM_{2.5}$ concentration should typically be lower. Despite this there were some days where $PM_{2.5}$ was higher than PM_{10} which is likely due to small discrepancies that can occur when different instruments are being used to monitor the different particulate size fractions³.

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 3 A particulate monitoring instrument with a PM₁₀ sampling head will collect particles 10 microns in size and smaller including particles 2.5 microns in size and smaller whereas an instrument with a $PM_{2.5}$ sampling head will only collect particles 2.5 microns in size and smaller. The BAM FH62 used for PM₁₀ has relative humidity control on the sampler inlet compared to the BAM 5014i used for PM_{2.5} at Tokoroa which has temperature control on the sampler inlet

Figure 3.12: Relationship between 24 hour average PM¹⁰ and PM2.5 concentrations measured in Tokoroa for winter 2017.

Tokoroa PM₁₀ concentrations from 2006 to 2017 are presented relative to air quality indicator categories in Figure 3.13. The proportion of concentrations within the "action" category was higher in 2017 at 3% compared to 2016 at 1.5% but is consistent with the typical range observed for previous years of 3% to 6%.

Figure 3.13: Comparison of 24 hour average PM¹⁰ concentrations measured in Tokoroa from 2006 to 2017 to air quality indicator categories.

The proportion of 24 hour average $PM_{2.5}$ concentrations within the "action" category for 2017 was 18% which is consistent with the 18% proportion of PM_{2.5} concentrations determined for 2016.

A comparison of monthly average PM_{10} concentrations for 2017 (Figure 3.14) shows a distinct seasonal pattern with a wintertime peak indicative of home heating impacts consistent with seasonal patterns observed for other airsheds in the Waikato region apart from Hamilton.

Figure 3.14: Monthly average PM¹⁰ concentrations measured in Tokoroa in 2017.

The number of days when the PM₁₀ standard of 50 μ g/m³ as a 24 hour average was exceeded, the maximum 24 hour average concentration and the $2nd$ highest 24 hour average concentration over the period 2006 to 2017 is shown in Figure 3.15. The NES compliance targets for Tokoroa are no more than three exceedances after 1 September 2016 and no more than one exceedance after 1 September 2020. It is unlikely that either target would be met in the absence of additional air quality management measures.

Figure 3.15: Number of days (left axis) when the PM¹⁰ standard of 50 µg/m³as a 24 hour average was exceeded compared with the maximum 24 hour average concentration and the 2nd highest 24 hour average concentration (right axis) measured from 2006 to 2017.

The annual average PM₁₀ concentration for Tokoroa for 2017 was 15.4 μ g/m³ which is towards the lower end of the range of annual average concentrations recorded at the site since 2006 (range 14.5 to 18.8 μ g/m³). The Ministry for the Environment specifies an annual average guideline for PM₁₀ of 20 μ g/m³ (MfE, 2002). The annual average PM_{2.5} concentration for Tokoroa for 2017 was 14.7 μ g/m³. This is higher than the WHO guideline for annual average PM_{2.5} of 10 μ g/m³. Refer to Appendix 2 for a full summary of PM₁₀ monitoring statistics for Tokoroa for the period 2006 to 2017 and PM2.5 monitoring statistics for the period 2015 to 2017.

A windrose showing the relative frequency and speed of winds from different directions for 2017, as measured at the Tokoroa station, is compared with the windrose for 2016 in Figure 3.16. The 2017 wind rose indicates two distinct prevailing wind directions, one from the westnorthwest and another from the southeast with a high frequency of low to moderate windspeeds which is very consistent with the windrose pattern observed for 2016. There is a lower frequency of calms (0.25%) recorded for 2017 compared to 4.57% in 2016.

Figure 3.16: Windrose of wind direction and windspeed data as measured at the Tokoroa station in 2016 and 2017.

Daily variations in meteorological conditions and hourly average PM $_{10}$ and PM $_{2.5}$ concentrations on the ten days when the 24 hour average PM₁₀ concentrations exceeded 50 μ g/m³ during 2017 are shown in Figure 3.17. The daily variations in PM_{10} and $PM_{2.5}$ concentrations presented in Figure 3.17 are typical of small urban towns in New Zealand where domestic home heating is the main contributor to elevated PM_{10} concentrations. The $PM_{2.5}$ diurnal pattern is generally a similar match to the PM₁₀ diurnal pattern, providing additional evidence that the PM₁₀ exceedances are being caused by combustion sources, but with a smoother profile. Meteorological conditions for Tokoroa include low wind speeds and air temperatures, a southeast wind direction during the evening/ overnight and morning and a west to northwest wind during the afternoon as indicated by the prevailing wind directions identified by the windrose in Figure 3.16.

Figure 3.17: Hourly average PM10, PM2.5, wind speed, wind direction and temperature on days when PM¹⁰ concentrations exceeded the NES in Tokoroa.

The Seasonal Mann Kendall test results for Tokoroa with the inclusion of 2017 data indicate a significant reduction in PM_{10} concentrations for the period 2009 to 2017 (MK-Stat of -2.3 and a

p-value of 0.02). For the period 2008 to 2017 the trend is not quite significant (MK-Stat of -1.7 and a p-value of 0.09) suggesting that more consistent reductions have occurred post 2009.

The analysis of air temperature over the period 2009 to 2017 also indicates a significant increase in air temperature over this period (MK-Stat of 1.95 and a p-value of 0.05).

A 32% reduction in wintertime PM₁₀ emissions was estimated to have occurred between 2012 and 2016 based on air emission inventory information (Wilton, 2016). The majority of emission reductions were as a result of reductions in emissions from domestic home heating (mainly woodburners and open fires).

Additional analysis has also been undertaken using methodologies which account for the impacts of varying meteorology in Tokoroa (Wilton, 2013a). Meteorological conditions conducive to high pollution in Tokoroa include days which have both low wind speeds and low temperatures over a specific period of the day, namely:

- Days with more than 15 hours when the hourly average wind speed was less than 2 metres per second; and
- Days where the average temperature from 8 pm to midnight is less than 5.36 degrees Celsius.

In summary, if these criteria are met then these are the days you would expect most exceedances to occur. Figure 3.18 provides a summary of the year-to-year variation of the proportion of high pollution potential days which resulted in exceedances in Tokoroa. This shows that there was a large decrease in the proportion of high pollution potential days with exceedances around 2013 and 2014. Since then, the proportion of high pollution potential days with exceedances has steadily increased again with 2017 producing a similar proportion to that identified in 2007 and 2008.

Figure 3.18: Year-to-year variation of the proportion of high pollution potential days which resulted in exceedances.

Figure 3.19 provides a summary of the year-to-year variation of the absolute number of high pollution potential days compared with the number of high pollution potential days that had exceedances. It shows that 2017 had a relatively high number of high pollution potential days compared to 2016 which had the lowest number of days qualifying as high pollution potential over the analysis period of 2006 to 2017.

Figure 3.19: Year-to-year variation of the absolute number of high pollution potential days compared with the total number of exceedance days.

If PM_{10} concentrations were decreasing, a consistent reduction in the proportion of high pollution potential days that resulted in breaches would be evident. Between 2006 and 2012 breaches occurred on 39-70% of days when meteorological conditions conducive to elevated PM₁₀ occurred. From 2013-2017 the proportion ranges from 22-41% but with an indication that the proportion of days with a breach has been increasing lately rather than decreasing. The reason for this is unclear.

Another way of comparing changes in air quality over time is to compare the average number of exceedances per year based on a five year rolling average (Figure 3.20). The use of a five year average removes some of the impact of the year to year variability in meteorological conditions whilst targeting the indicator of most concern, the number of exceedences of the 24 hour average PM₁₀ standard of 50 μ g/m³. Figure 3.20 provides some indication of a reduction in exceedences at Tokoroa with a consistent decrease in the five year rolling average exceedences from 2008-2012 (15 days) to 2013-2017 (8.8 days), noting that some of the reduction observed between in the 5 year average between 2015 and 2016 likely occurred as a result of favourable meteorological conditions during 2016.

Figure 3.20: Comparison of the five-year exceedance averages for the period 2006 to 2017.

It is clear from the data that despite reductions in PM_{10} concentrations in Tokoroa in recent years the interim NES target of no more than three exceedances by September 2016 has not been met and significant intervention would be required to meet the September 2020 target of no more than one exceedence per year. It has been estimated that an emission reduction of around 24% of 2007 emission levels is required to meet the 2016 target and around 43% to meet the 2020 target (Wilton, 2011).

3.5 PM¹⁰ monitoring in Taupo

There was one exceedance of the 24 hour average PM_{10} standard of 50 μ g/m³ for 2017. The maximum 24 hour average concentration was 52 μ g/m³ and was measured on 28 June 2017. There was no exceedance of the 24 hour average standard in 2016 with the last breach of the NES at this site occurring in 2013 when two exceedances of the 24 hour average standard (65 μg/m³ and 62 μg/m³) were measured. Figure 3.21 shows the 24 hour average concentrations of PM₁₀ for 2017 compared to the NES value of 50 μ g/m³.

Figure 3.21: 24 hour average PM¹⁰ concentrations measured at Taupo during 2017.

Figure 3.22 shows the changes in 24 hour average PM_{10} concentrations relative to air quality indicator categories at the Taupo site from 2006 to 2017. Data are adjusted for gravimetric equivalency only since 2007 so comparison of trends with pre 2006 data are limited⁴. There has been a gradual increase in the proportion of PM_{10} concentrations in the 'good' category since 2007.

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⁴ For 2006 the gravimetric partisol data from Gillies Ave were used in preference to the BAM primary school data and therefore it is considered appropriate to include the 2006 data for comparison with data collected after 2006 which was BAM data corrected against gravimetric data.

Figure 3.22: Comparison of 24 hour average PM¹⁰ concentrations measured at the Taupo site from 2006 to 2017 to air quality indicator categories.

A comparison of monthly average PM₁₀ concentrations for 2017 (Figure 3.23)⁵ shows a distinct seasonal pattern with a wintertime peak indicative of home heating impacts consistent with seasonal patterns observed for other airsheds in the Waikato region apart from Hamilton.

Figure 3.23: Monthly average PM¹⁰ concentrations measured in Taupo in 2017.

Figure 3.24 shows the number of days when the 24 hour average PM₁₀ standard of 50 μ g/m³ was exceeded, the maximum concentration and the 2nd highest concentration for 2006 to 2017. Data suggests improvements in PM₁₀ concentrations particularly since 2010.

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⁵ Due to missing data for March there was insufficient data to calculate a monthly average.

Figure 3.24: Number of days (left axis) when the 24 hour average standard of 50 µg/m³ was exceeded compared with the maximum concentration and the 2nd highest concentration (right axis) measured from 2006 to 2017.

The annual average PM₁₀ concentration for Taupo for 2017 was 13.5 μ g/m³ which is towards the lower end of the range of annual average concentrations recorded at the site since 2006 (range 11.9 to 20.4 μ g/m³). Refer to Appendix 3 for a full summary of PM₁₀ monitoring statistics for Taupo for the period 2001 to 2017.

A wind rose showing the relative frequency and speed of winds from different directions for 2017, as measured at the Taupo station, is compared with the windrose for 2016 in Figure 3.25. The 2017 wind rose indicates one distinct prevailing wind direction from the northeast and a second broader wind direction spanning the west to southwest quadrant. The windrose for Taupo also indicates a high frequency of low windspeeds (0.5 to 2.1 m/s) compared to windroses for Tokoroa and Hamilton.

Figure 3.25: Windrose of wind direction and windspeed data as measured at the Taupo station in 2016 and 2017.

The highest and second highest 24 hour average PM_{10} concentrations during 2017 were recorded on 28 June and 3 July respectively. Figure 3.26 shows variations in PM_{10} concentrations and meteorological data on these two days. Both graphs show a small morning peak in PM_{10} concentrations around 7 to 9 am and an increase in PM_{10} concentrations around 5pm with elevated concentrations for the duration of the evening. The variations are typical of those in small urban areas of New Zealand where emissions from solid fuel burning for domestic heating the dominant source of PM_{10} .

In Taupo, high PM₁₀ concentrations typically occur when the wind is from an easterly or south easterly direction and wind speeds are low. During the daytime the wind shifts to westerly, returning to east/south east during the evening (Wilton & Baynes, 2010). Meteorological conditions on 28 June and 3 July were reasonably consistent with these previously identified patterns.

Figure 3.26: Hourly average PM10, wind speed, wind direction and temperature on days when the 24 hour average PM¹⁰ concentration was elevated in Taupo.

Seasonal Mann Kendall test results (MK-Stat of -1.8 and p-value of 0.06) indicates that the improving trend identified in previous years is now less significant for the period 2008 to 2017 and is likely to be indicating a levelling out or slowing down of the previous significant improvement observed.

A 23% reduction in wintertime PM_{10} emissions was estimated to have occurred between 2009 and 2014 based on air emission inventory information (Wilton, 2015a). The majority of emission reductions were as a result of reductions in emissions from domestic home heating (mainly woodburners and open fires).

Meteorological conditions conducive to high pollution in Taupo include days which have both low wind speeds and low temperatures over a specific period of the day (Wilton 2013), namely:

- Days with more than 16 hours when the hourly average wind speed was less than 1 metres per second; and
- Days where the average temperature between 5 pm to midnight is less than 7.7 degrees Celsius.

In summary, if both criteria are met then these are the days you would expect exceedances to occur. Figure 3.27 provides a summary of the year-to-year variation of the proportion of high pollution potential days which resulted in exceedances in Taupo. This suggests a decrease in the proportion of high pollution potential days that resulted in exceedences of the 24 hour average standard of 50 µg/m³, with over 20% of high pollution potential days having exceedences between 2007 and 2009 compared with around 0% to 10% from 2010-2017.

Figure 3.27: Year-to-year variation in the proportion of high potential pollution days which resulted in exceedances.

Figure 3.28 provides a summary of the year-to-year variation of the absolute number of high pollution potential days compared with the number of high pollution potential days that had exceedances. This shows that for more recent years the number of high pollution potential days is still relatively high while the number of high pollution potential days resulting in exceedances has reduced. This is a strong indicator that PM_{10} concentrations have reduced in Taupo.

Figure 3.28: Year-to-year variation in the absolute number of high pollution potential days compared with the total number of exceedance days.

The 5 year average of exceedances per year from 2006 to 2017 is shown in Figure 3.29. Comparing averages over many years is another way of reducing some of the impact of year to year variability in meteorological conditions. There has been a consistent downwards trend in the 5 year exceedance average over the period 2006 to 2017.

Figure 3.29: Comparison of the five year exceedance averages for the period 2006 to 2017.

Based on these analyses and the current previous five year exceedance average of less than one, it is likely that Taupo will continue to meet the NES 2016 target of no more than one exceedance per 12 month period. But this will still be dependent on meteorological conditions i.e. a particularly cold and calm winter could still result in more than one exceedance within a 12 month period. Further reductions in emissions would provide a "safeguard" against future impacts of changes in meteorology.

3.6 NO² monitoring in Taupo

The results from NZTA's passive $NO₂$ monitoring programme for 2017 (refer to Figure 3.30) indicates that the annual average $NO₂$ concentration continues to remain well below the WHO annual average guideline of 40 μ g/m³. Seasonal Mann Kendall test results for monthly average NO² concentrations for Taupo for the period 2011 to 2017 indicate no statistically evident change. Note that the trend analysis was limited to 2011 to 2017 because monthly average data collected prior to 2011 has not been made available by NZTA.

Figure 3.30: Annual average NO² concentrations measured in Taupo (2007 to 2017).

3.7 PM¹⁰ monitoring in Te Kuiti

There were no exceedances of the 24 hour average PM_{10} standard of 50 μ g/m³ for 2017. Figure 3.31 shows 24 hour average PM_{10} concentrations measured in Te Kuiti during 2017. The maximum measured 24 hour average PM₁₀ concentration was 46 μ g/m³. No exceedances have been recorded in Te Kuiti since 2012 when one exceedance of 50 μ g/m³ (61 μ g/m³) occurred. The airshed has now been fully compliant with the NES for over five years and is no longer classified as "polluted" under the NES.

Figure 3.31: 24 hour average PM¹⁰ concentrations measured in Te Kuiti during 2017.

Figure 3.32 shows variations in the 24 hour average PM_{10} concentrations relative to air quality indicator categories in Te Kuiti from 2006 to 2017. A gradual increase in the proportion of PM_{10} concentrations in the 'good' category and decreases in the proportion of concentrations in the "action" and "alert" categories are an illustration of improvements in air quality in Te Kuiti.

Figure 3.32: Comparison of 24 hour average PM¹⁰ concentrations measured at the Te Kuiti site from 2006 to 2017 relative to air quality indicator categories.

A comparison of monthly average PM_{10} concentrations for 2017 (Figure 3.33) shows a very distinct seasonal pattern with a sharp wintertime peak indicative of home heating impacts consistent with seasonal patterns observed for other airsheds in the Waikato region apart from Hamilton.

Figure 3.33: Monthly average PM¹⁰ concentrations measured in Te Kuiti in 2017.

The number of days when the 24 hour average PM₁₀ standard of 50 μ g/m³ was exceeded, the maximum concentration and the 2^{nd} highest concentration from 2006 to 2017 are shown in Figure 3.34. Data is indicative of improvements in PM₁₀ concentrations since 2006.

Figure 3.34: Number of days (left axis) when the 24 hour average PM¹⁰ standard of 50 µg/m³ was exceeded compared with the maximum concentration and the 2nd highest concentration (right axis) measured from 2006 to 2017.

The annual average PM₁₀ concentration for 2017 was 14.1 μ g/m³ which is lower than the last 7 years where the annual average has ranged between 15 to 16 μ g/m³ and lower than the annual averages of approximately 18 μ g/m³ that were recorded prior to 2009. The Ministry for the Environment specifies an annual average guideline for PM₁₀ of 20 μ g/m³ (MfE, 2002). Refer to Appendix 4 for a full summary of PM_{10} monitoring statistics for Te Kuiti for the period 2003 to 2017.

A wind rose showing the relative frequency and speed of winds from different directions for 2017, as measured at the Te Kuiti station, is compared with the windrose for 2016 in Figure 3.35. The 2017 wind rose indicates one distinct prevailing wind direction from the northeast and a second broader wind direction spanning the northwest to southwest quadrant, similar to that indicated for Taupo. The 2017 windrose for Te Kuiti also indicates a higher frequency of high windspeeds compared with the windrose determined for 2016 but with a similar frequency of calms.

Figure 3.35: Windrose of wind direction and windspeed data as measured at the Te Kuiti station in 2016 and 2017.

Hourly variations in PM₁₀ concentrations and meteorological variables on 30 June and 26 July when the maximum and 2nd highest 24 hour average PM₁₀ concentrations of 46 μ g/m³ and 43 μ g/m³ respectively were measured are shown in Figure 3.36. On 30 June the hourly PM₁₀ profile is fairly typical of an urban PM_{10} pollution event with an elevated broad peak extending from the previous night and slowly reducing over the early hours of the morning followed by a peak in concentrations around 9 am, a drop off towards zero during the middle of the day and then a gradual increase again from about 5 pm. On the 26 July concentrations decrease after the previous night but the morning peak around 8 am is similar in magnitude to the later peak in PM₁₀ during the evening following the afternoon decrease.

Lower air temperatures were experienced on the 30 June (particularly in the morning) compared to the 26 July while wind speeds and patterns were similarly low throughout the day on both days.

Figure 3.36: Hourly average PM10, wind speed, wind direction, and temperature on 30 June and 26 July when the maximum 24 hour average PM¹⁰ concentrations of 46 µg/m³ and 43 µg/m³ respectively were recorded at Te Kuiti.

Seasonal Mann Kendall test results (MK-Stat of -2.6 and p-value of 0.008) provides strong evidence that PM_{10} concentrations in Te Kuiti have been decreasing over the period 2008 to 2017.

Another indicator of trends over time is the 5 year exceedance average (Figure 3.37) as averaging over a longer time period will remove some of the variation that occurs from year to year. A consistent downwards trend in the 5 year exceedance average over the period 2006 to 2017 is apparent.

Figure 3.37: Comparison of the five year exceedance averages for the period 2006 to 2017.

On a calendar year basis, Te Kuiti has had no more than one exceedance per year since 2011. However, two exceedances occurred in Te Kuiti over a 12 month period that straddled the period 2011 to 2012. The first exceedance of the breach occurred on 9 December 2011 and the second exceedance occurred on 2 July 2012.

An airshed ceases to be polluted when the PM $_{10}$ standard has not been breached in the airshed for five years. That is only one exceedance of the 24 hour average within any 12-month period during that five years. Therefore, the compliance period for Te Kuiti starts from 10 December 2011, the day after the first exceedance of the breach that occurred between 2011 and 2012. Because no exceedences of the 24 hour average standard of 50 μ g/m³ have occurred since 2012 the Te Kuiti airshed qualifies as a non-polluted airshed.

An air emission inventory undertaken in Te Kuiti in 2015 indicates that there has been a 14% reduction in PM₁₀ emissions since 2007 as a result of reductions in both domestic heating and industrial emissions (Wilton 2015b). While it appears emissions have reduced sufficient that further breaches may seem unlikely, it is possible that the NES for PM_{10} could still be breached in the airshed if particularly conducive meteorological conditions were experienced.

3.8 PM¹⁰ monitoring in Putaruru

There was one exceedance of the 24 hour average PM_{10} standard of 50 μ g/m³ for 2017. Figure 3.38 shows 24 hour average PM_{10} concentrations measured in Putaruru during 2017. The maximum measured 24 hour average PM₁₀ concentration was 55 μ g/m³ on 27 August 2017 with the next highest concentration of 50 μ g/m³ occurring on 16 June 2017. The last breach of the NES at this site occurred in 2014 when two exceedances of the 24 hour average standard of 50 μ g/m³ were recorded. Both of those exceedances occurred outside of the typical wintertime season in March and early April and had hourly profiles inconsistent with the typical diurnal pattern indicative of a home heating source. Prior to 2014 the previous NES breach occurred in 2009 when three exceedences of the 24 hour average PM₁₀ standard occurred.

Figure 3.38: 24 hour average PM¹⁰ concentrations measured at the Putaruru site during 2017.

Figure 3.39 compares variations in the 24 hour average PM_{10} concentrations relative to air quality indicator categories in Putaruru from 2007 to 2017. Data for 2017 indicates an improvement from previous years with less than 20% of concentrations being higher than a 24 hour average concentration of 16.5 μ g/m³ (33% of the NES standard).

Figure 3.39: Comparison of 24 hour average PM¹⁰ concentrations measured at the Putaruru site from 2007 to 2017 relative to air quality indicator categories.

An evaluation of monthly average PM₁₀ concentrations for 2017 (Figure 3.40)⁶ shows a distinct seasonal pattern with a wintertime peak indicative of home heating impacts consistent with seasonal patterns observed for other airsheds in the Waikato region apart from Hamilton.

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Doc # 12332239 Page 39 ⁶ Due to missing data for August and November there was insufficient data to calculate a monthly average.

Figure 3.40: Monthly average PM¹⁰ concentrations measured in Putaruru in 2017.

Figure 3.41 shows the number of days when the 24 hour average PM₁₀ standard of 50 μ g/m³ was exceeded, the maximum concentration and the 2nd highest concentration from 2007 to 2017. The 2006 data has been excluded from the comparison as monitoring only began half way through the winter season. The greatest number of exceedances and the highest PM_{10} concentrations occurred during 2008. However, it is worth noting that in 2008, two of the four recorded exceedances were in summer (February) and came about as a result of dust created by roadworks during the unusual drought conditions. The two exceedances that occurred during 2014 are also atypical and occurred out of season.

Data indicates no breaches of the NES for PM_{10} that occur as a result of typical winter air pollution since 2009. This suggests that in the absence of the atypical sources, Putaruru would be in compliance with the NES. However, because the atypical sources have not been identified it is possible that they could result in future breaches at the site.

Figure 3.41: Number of days when the 24 hour average standard of 50 µg/m³ was exceeded compared with the maximum concentration and the 2nd highest concentration measured from 2007 to 2017.

The annual average PM₁₀ concentration for Putaruru for 2017 was 11.6 μ g/m³ which is similar to the annual average of 11.8 μ g/m³ for 2016 and lower than previous annual averages recorded at the site. Refer to Appendix 5 for summary of PM_{10} monitoring statistics for Putaruru for the period 2006 to 2017.

Seasonal Mann Kendall test results for Putaruru (MK-Stat of -2.3 and p-value of 0.02) indicate that PM $_{10}$ concentrations have been improving over the period 2008-2017. This is consistent with the visual observation in Figure 3.39 of an improvement in PM_{10} in the "good" and "acceptable" air quality indicator categories over the period 2008 - 2012.

Another indicator of trends over time is the 5 year exceedance average (Figure 3.42) as averaging over a longer time period will remove some of the variation that occurs from year to year. A consistent downwards trend in the 5 year exceedance average over the period 2006 to 2017 is apparent.

Figure 3.42: Comparison of the five-year exceedance averages for the period 2006 to 2017.

An air emission inventory undertaken in Putaruru in 2015 indicates that there has been a 62% reduction in PM₁₀ emissions since 2006 as a result of reductions in both domestic heating and industrial emissions (Wilton 2015b).

The overall indication for Putaruru is that exceedance numbers are tracking downwards and that Putaruru will continue to meet the NES 2016 target of no more than one exceedance per 12 month period. However, this will still be dependent on meteorological conditions i.e. a particularly cold and calm winter could still result in more than one exceedance within a 12 month period. Further reductions in emissions would provide a "safeguard" against future impacts of changes in meteorology.

3.9 PM¹⁰ monitoring in Turangi

There were no exceedances of the 24 hour average PM_{10} standard of 50 μ g/m³ for 2017. Figure 3.43 shows 24 hour average PM_{10} concentrations measured in Turangi during 2017. The maximum 24 hour average PM₁₀ concentration of 28 μ g/m³ was recorded on both 14 and 17 May 2017. The maximum 24 hour average PM₁₀ concentration recorded at this site was 41 μ g/m³ which was recorded in June 2015.

Figure 3.43: Daily average PM¹⁰ concentrations measured at the Turangi site for 2017.

Figure 3.44 shows 24 hour average concentrations of PM_{10} relative to air quality indicator categories at Turangi from 2009 to 2017. In 2017, as for previous years, approximately 90% of days experienced 24 hour average PM_{10} concentrations within the "good" category. On all other days PM_{10} was within the "acceptable" category.

Figure 3.44: Comparison of 24 hour average PM¹⁰ concentrations measured at the Turangi site from 2009 to 2017 relative to air quality indicator categories.

Figure 3.45 shows that there have been no exceedances of the 24 hour average PM $_{10}$ standard of 50 μ g/m³ over the monitoring period 2009 to 2017. While the maximum 24 hour average PM₁₀ concentration measured at the site of 41 μ g/m³ was recorded recently in 2015, a visual comparison of the data (Figure 3.40 and 3.41) does not indicate a worsening trend.

Figure 3.45: Comparison of the maximum 24 hour average PM¹⁰ concentration and the 2nd highest average PM¹⁰ concentration measured from 2009 to 2017.

The annual average PM₁₀ concentration for Turangi for 2017 was 9 μ g/m³ and is consistent with previous years. The Ministry for the Environment specifies an annual average guideline for PM_{10} of 20 μ g/m³ (MfE, 2002). Refer to Appendix 6 for a full summary of PM₁₀ monitoring statistics for Turangi for the period 2009 to 2016.

3.10 NO² monitoring in Cambridge

The results from NZTA's passive $NO₂$ monitoring programme for 2017 (refer to Figure 3.46) indicate that the annual average $NO₂$ concentration continues to remain below the WHO annual average guideline of 40 μ g/m³. Seasonal Mann Kendall test results for monthly average NO₂ concentrations for Cambridge for the period 2011 to 2017 indicate no statistically evident change. Note that the trend analysis was limited to 2011 to 2017 because monthly average data collected prior to 2011 has not been made available by NZTA.

Figure 3.46: Annual average NO² measured in Cambridge (2007 to 2017).

3.11 NO² monitoring in Te Awamutu

The results from NZTA's passive $NO₂$ monitoring programme for 2017 (refer to Figure 3.47) indicates that the annual average $NO₂$ concentration continues to remain below the WHO annual average guideline of 40 μ g/m³. Seasonal Mann Kendall test results for monthly average NO₂ concentrations for Te Awamutu for the period 2011 to 2017 indicate no statistically evident change. Note that the trend analysis was limited to 2011 to 2017 because monthly average data collected prior to 2011 has not been made available by NZTA.

Figure 3.47: Annual average NO² measured in Te Awamutu (2007 to 2017).

3.12 PM¹⁰ monitoring in Morrinsville

There were no exceedances of the 24 hour average PM_{10} standard of 50 μ g/m³ for 2017. Figure 3.48 shows 24 hour average PM_{10} concentrations measured in Morrinsville during 2017. The maximum 24 hour average PM₁₀ concentration was 35 μ g/m³ and was measured on 31 July. The second highest 24 hour average PM₁₀ concentration of 33 μ g/m³ was measured on 27 June. In 2015 a maximum 24 hour average PM₁₀ concentration of 45 μ g/m³ was measured during July. No exceedances of the 24 hour average PM₁₀ standard of 50 μ g/m³ have been recorded at the site since monitoring commenced in 2015.

Figure 3.48: 24 hour average PM¹⁰ concentrations measured at Morrinsville during 2017.

A comparison of 24 hour average concentrations of PM_{10} relative to air quality indicator categories at Morrinsville for 2017 indicates around 88% of days experienced PM_{10} concentrations within the 'good' category, 11% within the acceptable category and 1% within the alert category, similar to 2016.

The annual average PM₁₀ concentration for 2017 for Morrinsville was 11.3 μ g/m³ and is at the low end of the range of annual averages determined for airsheds in the Waikato region. Refer to Appendix 7 for a full summary of PM₁₀ monitoring statistics for Morrinsville for 2015 to 2017.

3.13 PM¹⁰ monitoring in Thames

There were no exceedances of the 24 hour average PM₁₀ standard of 50 μ g/m³ for 2017. Figure 3.49 shows 24 hour average PM_{10} concentrations measured in Thames during 2017. The maximum 24 hour average PM₁₀ concentration was 29 μ g/m³ and was measured on 1 July compared with the maximum of 34 μ g/m³ measured for 2016.

Figure 3.49: Daily average PM¹⁰ concentrations measured at Thames during 2017.

A comparison of 24 hour average concentrations of PM_{10} relative to air quality indicator categories at Thames for 2017 indicates around 93% of days experienced PM_{10} concentrations within the 'good' category and 7% within the acceptable category.

The annual average PM₁₀ concentration for 2017 for Thames was 8.5 μ g/m³. Refer to Appendix 8 for a full summary of PM₁₀ monitoring statistics for Thames for 2016 and 2017.

4 Summary and conclusions

Monitoring of PM¹⁰ was undertaken in Hamilton, Tokoroa, Taupo, Te Kuiti, Putaruru, Turangi, Morrinsville and Thames during 2017. In addition to this, PM_{2.5} was monitored in Tokoroa. This monitoring was targeted mainly at identifying the impacts from domestic home heating sources.

Monitoring of NO₂ was undertaken by NZTA with contributory funding from WRC in Hamilton, Cambridge, Te Awamutu and Taupo for purposes of identifying the impacts from traffic related emissions. Monitoring of benzene, toluene, xylenes and ethylbenzene (BTEX) were also undertaken in Hamilton as additional indicators of traffic related emissions.

In 2017 the 24 hour average standard for PM_{10} of 50 μ g/m³ was exceeded 10 times in Tokoroa. One exceedance each was also recorded in Taupo and Putaruru but there was no breach of the NES as the regulations allow for one exceedance per 12 month period. Annual average concentrations of PM₁₀ were also all within the annual average guideline of 20 μ g/m³ in all airsheds.

Table 4.1 summarises the maximum 24 hour average PM_{10} concentration recorded, the number of exceedances of the 24 hour average PM₁₀ standard of 50 μ g/m³ and the annual average PM₁₀ concentration in all airsheds monitored in 2017.

Table 4.1: Summary of PM¹⁰ monitoring results for 2017.

Table 4.2 shows the average number of exceedences of the 24 hour average PM_{10} standard at Hamilton, Tokoroa, Taupo, Putaruru and Te Kuiti over a five-year period and the trend in monthly average PM₁₀ concentrations determined by Mann Kendall analysis. Based on historical monitoring, these five airsheds have the greatest potential for particulate pollution.

The trend analysis indicates that PM_{10} concentrations are decreasing in Te Kuiti, Tokoroa, and Putaruru. A previous decreasing trend identified for Taupo is no longer significant and indicates that concentrations are starting to level out. No change has been identified for the two Hamilton monitoring stations based at Ohaupo Road and Claudelands Event Centre grounds.

A comparison of the changes in the five year exceedance average for PM_{10} for these five airsheds is also provided in Figure 4.1 which provides additional evidence of the improvements observed over the long term for these airsheds.

Figure 4.1: Comparison of changes in the annual exceedance average (5 year rolling average) for Hamilton, Tokoroa, Taupo, Te Kuiti and Putaruru.

During 2017 the 24 hour average $PM_{2.5}$ concentration in Tokoroa exceeded the WHO guideline of 25 μ g/m³ on 53 occasions. The annual average PM_{2.5} concentration for Tokoroa was 14.7 μ g/m³ which also exceeds the WHO annual average guideline of 10 μ g/m³.

The results from passive NO₂ monitoring indicate that the Lorne Street/Ohaupo Road site and Greenwood Road/Killarney Road site in Hamilton continue to exceed the World Health Organisation (WHO) annual guideline of 40 μ g/m³ for NO₂. No exceedances of the WHO annual guideline for $NO₂$ at the other Hamilton sites and the Cambridge. Te Awamutu and Taupo sites have been identified. Long term seasonal trend analysis provides no indication of either a worsening or improving trend for monthly average $NO₂$ concentrations at any of the monitoring sites for the period 2011 to 2017.

The highest annual average benzene concentration of 2.2 μ g/m³ was measured at the Greenwood Street monitoring site in Hamilton and is lower than the Ministry for the Environment's annual guideline of 3.6 μ g/m³. Concentrations of toluene, xylene and ethylbenzene were also well within acceptable levels at all sites. An improving or "levelling" trend is evident for annual average concentrations of benzene at all sites with the exception of Peachgrove Road which shows an increase for 2016 and 2017 compared with the previous 3 to 4 years. A new supermarket opening on this road, near to the monitoring site location at the end of 2015 is the likely cause of the increase observed.

A total of 70% of gazetted airsheds in the Waikato region have now been monitored for PM₁₀. Five of these airsheds have been identified as requiring ongoing monitoring based on the identification of previous PM_{10} exceedances, namely Tokoroa, Taupo, Te Kuiti, Putaruru and

Hamilton. Improving trends have been identified in Tokoroa, Taupo, Te Kuiti and Putaruru which have mainly been attributed to reductions in emissions from home heating sources. The Tokoroa airshed, however, is unlikely to meet the NESAQ compliance target by September 2020 without additional or alternative intervention. Monitoring of PM_{2.5} over the last three years in Tokoroa indicates that compliance with the WHO annual average guideline, if adopted as a national standard, is also unlikely to be met.

While woodburners for domestic home heating have been identified as the main source of poor air quality, traffic sources in Hamilton have been identified as significant contributors to poor air quality in localised areas close to busy traffic routes and intersections based on the passive $NO₂$ monitoring data.

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Summary of PM¹⁰ concentrations measured at Ohaupo Road station in Hamilton for 2012 to 2017.

Summary of PM¹⁰ concentrations measured at Claudelands Event Centre station in Hamilton for 2014 to 2017.

1. Winter average rather than annual average due to limited dataset.

Summary of PM¹⁰ concentrations measured using a TEOM at Peachgrove Road station in Hamilton from 2006 to 2013 (gravimetric correction of data applied¹).

1. Corrected PM_{10} data = 1.19975 x RawTEOMdata - 3.9182.

2. Winter average rather than annual average due to limited dataset.

1. Gravimetric correction of data applied from 2006 onwards.

1. Gravimetric correction of data applied from 2006 onwards.

Summary of PM2.5 concentrations measured at the Tokoroa monitoring station from 2015 to 2017

Summary of PM¹⁰ concentrations measured at the Taupo monitoring station from 2001 to 2008¹ .

Summary of PM₁₀ concentrations measured at the Taupo monitoring station from 2009 to 2017¹.

1. 2007 - 2008 data have been updated from that reported in the 2007 and 2008 reports based on a more recent (2009) adjustment factor. 2006 data were gravimetric at Gillies Ave. Data post 2006 has been adjusted for gravimetric equivalency. Note the 2008 monitoring report used a different equation and reported six exceedances of 50 μ g m⁻³ for 2007 compared with three exceedances reported here.

2. To avoid seasonal bias in missing data, annual averages for gravimetric data collected prior to 2007 have been calculated based on the average of the individual seasonal averages (i.e. Jan to Apr, May to Aug, Sep to Dec). Annual average calculations from 2007 onwards have been based on averaging of all data (it makes no difference which method is used because there is very little missing data).

3. For the years 2001 to 2006 both measured and reported exceedences are shown (in brackets). Reported exceedences are a statistical extrapolation of measured exceedences after accounting for non sample days.

Summary of PM¹⁰ concentrations measured at the Te Kuiti monitoring station from 2003 to 2009.

Summary of PM¹⁰ concentrations measured at the Te Kuiti monitoring station from 2010 to 2017.

Summary of PM¹⁰ concentrations measured at the Putaruru monitoring station 2017.

Appendix 7

Summary of PM¹⁰ concentrations measured at the Morrinsville monitoring station from 2015 to 2017.

Summary of PM¹⁰ concentrations measured at the Thames monitoring station for 2016 and 2017.